

Evolution of dispersal and life history interact to drive accelerating spread of an invasive species

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1 ABSTRACT

2 Populations on the edge of an expanding range are subject to unique evolutionary pressures
3 acting on their life-history and dispersal traits. Empirical evidence and theory suggest that
4 traits there can evolve rapidly enough to interact with ecological dynamics, potentially giving
5 rise to accelerating spread. Nevertheless, which of several evolutionary mechanisms drive this
6 interaction between evolution and spread remains an open question. We propose an integrated
7 theoretical framework for partitioning the contributions of different evolutionary mechanisms to
8 accelerating spread, and we apply this model to invasive cane toads in northern Australia. In
9 doing so, we identify a previously unrecognized evolutionary process that involves an interaction
10 between life-history and dispersal evolution during range shift. In roughly equal parts, life-
11 history evolution, dispersal evolution, and their interaction led to a doubling of distance spread
12 by cane toads in our model, highlighting the potential importance of multiple evolutionary
13 processes in the dynamics of range expansion.

14 INTRODUCTION

15 There is a growing appreciation that rapid evolution can alter ecological processes in important
16 ways (e.g., Ellner *et al.*, 2011; Schoener, 2011). Rapid adaptation, for example, can change
17 population dynamics in managed fisheries (Law, 2000) and fundamentally alter disease dynamics
18 (Ewald, 1994). The rate at which a species shifts its range is another ecological process for which
19 rapid evolution is increasingly seen as important. Predicting a species' rate of range shift has
20 long been a major focus of ecologists (e.g., Elton, 1958), because the magnitude of damage
21 caused by noxious invasive species is determined by the speed and extent to which these species
22 spread (Epanchin-Niell & Hastings, 2010). Range shift is also critical to our understanding of
23 how species track climate change: both historical climate change (Skellam, 1951; Clark, 1998)
24 as well as the rapid anthropogenic shifts we are currently experiencing (Parmesan & Yohe,
25 2003). Thus, a full understanding of the drivers of range expansion is an imperative for applied
26 ecological research, and rapid evolution may well be an important facet of that understanding.

27 A phenomenon of interest in any range expansion is the possibility of increasing spread rates
28 over time. Accelerating spread has been observed in many invasions (Veit & Lewis, 1996; Crooks
29 & Soule, 1999; Urban *et al.*, 2008), and four purely ecological mechanisms—long-distance dis-
30 persal, Allee effects, density-dependent dispersal, and temporal variability—have been invoked
31 to account for this phenomenon (Kot *et al.*, 1996; Veit & Lewis, 1996; Shigesada & Kawasaki,
32 1997; Ellner & Schreiber, 2012). An additional, evolutionary, explanation has also been pro-
33 posed. This evolutionary explanation builds on the well-established principle that spread rate is
34 determined jointly by population growth and dispersal at the invasion front (Fisher, 1937) and
35 argues that evolved changes in either dispersal or the life-history traits governing population
36 growth can cause accelerating spread (Holt *et al.*, 2005).

37 We might expect life-history and dispersal traits to evolve during spread for three related
38 reasons. First, dispersal phenotypes are spatially assorted on the invasion front—only the best
39 dispersers are represented at the leading edge of the front—and this leads to assortative mating
40 by dispersal (Shine *et al.*, 2011). Second, because the leading edge of the invasion front is

41 at low density relative to populations behind it, individuals at the edge may experience little
42 competition from conspecifics and so have high absolute fitness relative to individuals in the core
43 of the range. Together, spatial assortment and the increased fitness of invasion-front individuals
44 drive the evolution of increasing dispersal on the leading edge of the invasion front: an interaction
45 of evolutionary forces that has been termed “spatial selection” (Phillips *et al.*, 2008b, 2010b).
46 Third, life-history traits of populations at the edge can evolve in response to natural selection
47 in the traditional sense. That is, life-history phenotypes that result in greater contributions to
48 population growth under low-density conditions at the invasion front will be represented more
49 among future occupants of the invasion front, wherever that may be (Holt *et al.*, 2005; Phillips,
50 2009; Perkins, 2012).

51 Recent theoretical work has shown that these evolutionary forces can all generate phenotype
52 change at the edge of an expanding range, that this change gives rise to accelerating spread,
53 and that the extent of spread acceleration observed in nature can be achieved with modest
54 heritabilities and realistic levels of selection (Travis & Dytham, 2002; Travis *et al.*, 2005, 2009;
55 Burton *et al.*, 2010; Bénichou *et al.*, 2012; Bouin *et al.*, 2012; Perkins, 2012). This theoretical
56 work has meanwhile been corroborated by numerous empirical demonstrations of evolved phe-
57 notypic changes in range-shifting populations (Simmons & Thomas, 2004; Hughes *et al.*, 2007;
58 Léotard *et al.*, 2009; Rogers & Siemann, 2004; Phillips *et al.*, 2006, 2008a; Phillips, 2009) and
59 the implication that such phenotype changes have resulted in accelerating spread in at least one
60 case (Phillips *et al.*, 2008a). Despite these multiple lines of evidence, an empirical link between
61 trait evolution and accelerating spread has yet to be made.

62 This linkage is necessary because it is currently not known the extent to which accelerating
63 spread is derived from rapid evolution in real-world examples. Further, we currently have no idea
64 as to the relative roles of evolution of traits with direct impacts on population growth (hereafter
65 “life-history traits”) versus those directly related to dispersal (hereafter “dispersal traits”) in
66 driving accelerating spread. Analyses of stage-structured integrodifference equations (Neubert
67 & Caswell, 2000; Caswell *et al.*, 2003; Jongejans *et al.*, 2008) and spatial integral projection

68 models (Jongejans *et al.*, 2011) show that shifts in either life-history or dispersal traits can
69 impact spread rates, but these analyses do not account for the evolutionary processes that may
70 or may not have led to those differences.

71 To integrate the three evolutionary forces outlined above and quantify their potential contri-
72 butions to spread, we propose a general model that combines the well-known theories of stage-
73 structured population dynamics, integrodifference equations, and evolutionary quantitative ge-
74 netics. The structure of this model is necessarily more complex than other quantitative-genetic
75 models employed for the study of spatial phenomena in evolutionary ecology (e.g., Kirkpatrick
76 & Barton, 1997), because those models do not have the capacity for dispersal evolution (al-
77 though see Bénichou *et al.*, 2012; Bouin *et al.*, 2012). At the same time, the model we propose is
78 more easily parameterized with data and makes less specific assumptions than individual-based
79 models (e.g., Burton *et al.*, 2010). The resulting model is complex but can nonetheless be param-
80 eterized using commonly measured data: life tables, dispersal kernels, phenotype distributions,
81 and heritabilities. Here we apply this model to the well-studied cane toad invasion of Australia,
82 for which data to parameterize the model are readily available.

83 The cane toad (*Rhinella marina*) was first introduced in northeastern Australia in 1935.
84 Since then it has spread to occupy more than 1.3 million km² of the continent (Urban *et al.*,
85 2007). Importantly, while toads initially spread at around 10 km·yr⁻¹, their spread rate across
86 northern Australia has steadily increased so that they now spread at around 50 km·yr⁻¹ (Urban
87 *et al.*, 2008; Phillips *et al.*, 2006). Furthermore, life-history evolution and dispersal evolution
88 have both been documented in the toads' invasion across northern Australia (Phillips *et al.*,
89 2006, 2008a, 2010b; Phillips, 2009), and contemporary estimates of the phenotype distributions
90 of two life-history traits (tadpole stage duration, metamorph growth) and one dispersal trait
91 (dispersal distance) are available from near the introduction site (Gordonvale) and at a more
92 recently invaded population (Timber Creek).

93 The confluence of these observations affords us a unique opportunity to place our analysis
94 of the relative contributions of different evolutionary processes to spatial spread in an empir-

95 ical context. Initializing the model with phenotypes from the introduction site, we compare
96 phenotype evolution and distance spread after 72 years under model scenarios with differing
97 assumptions about trait heritability. Examining differences in phenotype distributions and dis-
98 tance spread between model scenarios allows us to isolate the relative effects of the evolution
99 of life-history versus dispersal traits. Under our model, evolutionarily-derived increases in cane
100 toad spread are driven roughly equally by life-history evolution, dispersal evolution, and, unex-
101 pectedly, by an interaction between the two. Combined, these evolutionary processes more than
102 double distance spread in the model.

103 **METHODS**

104 **Methodological Overview**

105 We developed a model with the specific intent of examining distance spread under different
106 evolutionary scenarios. Our objective was not to determine how close different evolutionary
107 scenarios came to emulating the observed spread of toads, but to understand the relative impacts
108 of different evolutionary mechanisms on the joint dynamics of phenotype change and spatial
109 spread. We realized different evolutionary scenarios by setting initial heritabilities either to zero
110 or to some realistic positive value. The model was parameterized wherever possible with data,
111 and initial heritabilities were selected such that the model produced changes in phenotype means
112 close to differences between contemporary estimates of phenotype means in populations located
113 near the introduction site (Gordonvale) and at a more recently invaded area (Timber Creek).

114 **The Variables**

115 The model we use is designed to describe the dynamics of population densities, continuous
116 genotype distributions, and continuous phenotype distributions of multiple life stages in discrete
117 time and continuous space. A general framework for modeling such dynamics for any collection
118 of traits and an arbitrary categorization of life stages is presented in the Supporting Information.
119 The following description pertains to an application of that framework to cane toads.

120 Population densities of cane toad juveniles $J_t(x)$ and adults $A_t(x)$ are censused once per
 121 year t along a single spatial dimension x . These densities are further distributed across values
 122 of three quantitative traits: duration of the tadpole stage (T), metamorph growth rate (M),
 123 and dispersal tendency (D). Due to how they enter the model, we sometimes refer to T and M
 124 together as “life-history traits” (L) and to D as the “dispersal trait.”

125 Also at each time t and location x , the model follows the joint probability distributions,
 126 $\psi_{J,t}(\mathbf{g}, x)$ and $\psi_{A,t}(\mathbf{g}, x)$, of genotypes, $\mathbf{g} = (g_T, g_M, g_D)$, defined on the same scales of measure-
 127 ment as the traits themselves. At the onset of an invasion, we assume that these genotypes
 128 are multivariate normally distributed about their means with additive genetic variance \mathbf{G} . The
 129 bearers of each genotype combination possess phenotypes, $\mathbf{z} = (z_T, z_M, z_D)$, that are multivari-
 130 ate normally distributed with mean equal to their genotype value and variance \mathbf{E} attributable
 131 to various environmental and non-additive genetic sources. Consequently, joint phenotype dis-
 132 tributions, $\psi_{J,t}(\mathbf{z}, x)$ and $\psi_{A,t}(\mathbf{z}, x)$, are also multivariate normally distributed with variance
 133 $\mathbf{P} = \mathbf{G} + \mathbf{E}$ at the onset of invasion. Thereafter, all genotype and phenotype distributions are
 134 free to depart from their initial forms due to selection and gene flow. The initial heritability
 135 of trait i , h_i^2 , is then equal to $\mathbf{G}_{i,i}/\mathbf{P}_{i,i}$. We constrained all covariances between traits to zero,
 136 although the modeling framework is flexible in this regard.

137 **The Model**

138 We now describe the biological assumptions and processes governing the spatiotemporal dynam-
 139 ics of the aforementioned variables for cane toads. A detailed mathematical exposition of the
 140 general modeling framework and additional information about the numerical implementation of
 141 the cane toad model are available in the Supporting Information.

142 For the population-dynamic component of the model, we apply an existing model of cane toad
 143 population dynamics (Lampo & De Leo, 1998) at local populations along the spatial dimension
 144 x . Per-capita recruitment to the juvenile stage is a product of clutch size ϕ , egg survival σ_E ,
 145 tadpole survival σ_T , and metamorph survival σ_M (Fig. 1, top left). Survivals through the

146 tadpole and metamorph stages are not fixed, however (Fig. 1, top). Assuming constant daily
 147 mortality risks for tadpoles and metamorphs, survival through each of these stages depends on
 148 how long it takes for toads to progress through them. Accordingly, maximal survival through
 149 the tadpole stage under ideal conditions can be written as

$$\hat{\sigma}_T(z_T) = \hat{\sigma}_{T,\text{daily}}^{z_T}, \quad (1)$$

150 where $\hat{\sigma}_{T,\text{daily}}$ is maximal daily survival. Consistent with Lampo & De Leo (1998), we define
 151 realized survival through the tadpole stage as a density-dependent function

$$\sigma_T(z_T) = \frac{\hat{\sigma}_T(z_T)}{1 + dT}, \quad (2)$$

152 where d is the strength of density dependence and $T = \phi\sigma_E A$ is tadpole density. The metamorph
 153 phenotype, on the other hand, is defined as a growth rate. By stipulating that passage through
 154 the metamorph stage requires some critical amount of growth, we can calculate the survival of
 155 a metamorph with a relatively fast or slow growth rate by raising baseline metamorph survival
 156 $\tilde{\sigma}_M$ to a power of the ratio of the baseline growth rate \tilde{z}_M with that individual's growth rate
 157 z_M ; i.e.,

$$\sigma_M(z_M) = \tilde{\sigma}_M^{\tilde{z}_M/z_M}. \quad (3)$$

158 After surviving those preliminary stages, a proportion σ_J of juveniles survives to adulthood in
 159 the second year (Fig. 1, right), and a proportion σ_A of adults survives to each year thereafter
 160 (Fig. 1, bottom).

161 The dynamics of local populations distributed along the spatial dimension x are linked by
 162 a dispersal event each year. We assume that juveniles do not move far enough to be of any
 163 consequence at a geographic scale, so we only model dispersal explicitly for adults. The outcome
 164 of this dispersal event is governed by a probability density function $k(x, x')$ that determines how a
 165 population at each location x' redistributes itself to all other locations x . We implement dispersal

166 with a kernel that depends on the distance $|x - x'|$ between locations and on an individual's
 167 dispersal phenotype z_D (Fig. 1, bottom).

168 Phenotype-dependent survival and dispersal in the model allow for the possibility of geneti-
 169 cally based evolution, provided that there is variation among genotypes (i.e., any entry of $\mathbf{G} > 0$).
 170 Because our model specifies the relationship between the distributions of phenotypes and the
 171 genotypes that underlie them, we can directly calculate how phenotype-dependent survival and
 172 dispersal shape the genotype distributions, $\psi_{J,t}(\mathbf{g}, x)$ and $\psi_{A,t}(\mathbf{g}, x)$. Following selection on ju-
 173 veniles, maturation to adulthood, and adult dispersal, adult toads mate randomly to produce
 174 the distribution of genotypes that will enter the juvenile cohort in the following year (Fig. 1,
 175 left). The full mathematical form of the model can be reproduced by applying the preceding
 176 details to the general model formulation presented in the Supporting Information.

177 **Parameterization**

178 Distributions of life-history phenotypes in our model were parameterized by fitting normal dis-
 179 tributions to the relevant subset of the values of z_T and z_M measured in Phillips (2009) (Fig. 2,
 180 bottom left and center). Specifically, we used phenotype measurements from juveniles reared in
 181 a common laboratory environment whose parents were collected *in situ* near either Gordonvale
 182 or Timber Creek and then bred in a common laboratory environment (Phillips, 2009). Thus,
 183 there is strong evidence that the differences between the distributions of z_T and z_M between
 184 these populations are genetic in origin. The fitted distributions from offspring of toads collected
 185 near Gordonvale served as the initial conditions of these distributions in our model at the onset
 186 of invasion. Likewise, the fitted distributions from offspring of toads collected near Timber Creek
 187 served as the empirical benchmark against which the life-history phenotypes in our model were
 188 compared after 72 years of spread. Values of all other parameters in the population-dynamic
 189 component of the model were taken from Lampo & De Leo (1998).

190 The dispersal component of the model was parameterized using radiotracking data from
 191 Phillips *et al.* (2008a). The subset of toads from that study that we used here were collected *in*

192 *situ* near either Gordonvale or Timber Creek and held in a common environment for two months.
193 Thereafter, toads were released at a common location in the field and radiotracked for five nights,
194 rendering a list of daily movement distances and turning angles for each released individual.
195 This experimental design ensured that environmental effects at the time of observation were
196 controlled for, and a subsequent study (Phillips *et al.*, 2010a) demonstrated a clear genetic
197 basis to differences between the populations at Gordonvale and Timber Creek. To then obtain
198 a distribution of displacements over 180 days (the approximate length of the active season for
199 toads in northern Australia), we summed 180 resampled daily movements for each toad 10^5 times.
200 Assuming that daily movements are independent and identically distributed for each individual,
201 the sum of daily movements over the course of a season should converge to a normal distribution.
202 With the further assumption that movements were isotropic, and thus that there was zero mean
203 displacement, each individual's normal dispersal kernel has a single variance parameter σ^2 that
204 determines the scale of its dispersal. Finding that these individual variances are approximately
205 lognormally distributed, we defined the dispersal phenotype as $z_D = \log(\sigma^2)$ to obtain a normally
206 distributed dispersal phenotype appropriate for our model. The distribution of z_D at the onset of
207 invasion was then parameterized by fitting a normal distribution to all $\log(\sigma^2)$ of toads collected
208 near Gordonvale (Fig. 2, bottom right). Likewise, the normal distribution of z_D fitted to all
209 $\log(\sigma^2)$ of toads collected near Timber Creek served as the empirical benchmark against which
210 the dispersal phenotype in our model was compared after 72 years of spread (Fig. 2, bottom
211 right).

212 The evolutionary component of our model was parameterized by setting the initial heri-
213 tability of each trait to the value that resulted in a change in phenotype mean after 72 years
214 equal to the difference between phenotype means measured at Gordonvale and Timber Creek
215 (Fig. 2, bottom). For the life-history traits, we picked initial heritabilities that satisfied this
216 criterion given a simplified model of phenotype change in which selection at the invasion front
217 was assumed to remain consistent for the very small population size that defined the location of
218 the moving invasion front (details available in the Supporting Information). Because no simpler

219 model of dispersal evolution in this context is possible, we used the full model to pick an initial
 220 heritability for the dispersal trait consistent with the difference between phenotype means at
 221 Timber Creek and Gordonvale. This resulted in the following values for the initial heritabilities:
 222 $h_T^2 = 0.10$, $h_M^2 = 0.16$, and $h_D^2 = 0.21$.

223 Analysis

224 Our primary aim was to determine the relative contributions of natural selection (on life-history)
 225 versus spatial selection (on dispersal) as drivers of phenotypic change and spread acceleration
 226 on moving invasion fronts. To determine the consequences of different scenarios in which these
 227 factors play varying roles, we compared phenotype means at the invasion front and distances
 228 spread after 72 years under different scenarios. The four scenarios we examined allowed neither,
 229 either, or both sets of traits to evolve by manipulating which traits had a positive initial heri-
 230 tability. Hereafter, we will refer to these scenarios using the convention L_0D_0 , L_+D_0 , L_0D_+ , and
 231 L_+D_+ , where the subscripts of L denote the positivity of both h_T^2 and h_M^2 , and the subscripts
 232 of D denote the positivity of h_D^2 . Positive values of initial heritabilities were consistent across
 233 scenarios.

234 RESULTS

235 Allowing traits to evolve led to substantial improvements in the model's ability to reproduce the
 236 phenomenon of accelerating spread observed by Urban *et al.* (2008). In the absence of genetic
 237 variation and evolution (L_0D_0), spread in the model proceeded at a constant rate to a distance
 238 of 470 km after 72 years (Fig. 3). Thus, the best available data indicate that variable but non-
 239 evolving dispersal is not sufficient to account for the accelerating spread of toads. Evolution
 240 of life-history and dispersal traits, run separately, increased distance spread to 601 km (L_+D_0)
 241 and 677 km (L_0D_+), respectively. When life-history and dispersal traits were allowed to evolve
 242 simultaneously (L_+D_+), spread accelerated markedly to a distance of 1004 km (Fig. 3): a 114%
 243 increase in distance spread relative to the scenario with no evolution.

244 Another important, and more general, result to emerge is that the extent to which traits
245 evolve in the model is not consistent across scenarios. In particular, when all phenotypes were
246 allowed to evolve independently in scenario L_+D_+ , the life-history traits fell slightly short of
247 their final values achieved in the L_+D_0 scenario, but the dispersal trait greatly exceeded the
248 value it achieved in the L_0D_+ scenario (Fig. 2). The fact that the dispersal trait changed much
249 more in L_+D_+ than in L_0D_+ implies an interaction between life-history and dispersal evolution
250 during spread. Such an interaction can be explained by an increase in the strength of spatial
251 selection acting on the dispersal trait. Because both spatial sorting and natural selection interact
252 to drive the evolution of dispersal during spread (together, spatial selection), when population
253 growth increases due to natural selection the rate of dispersal evolution also increases. This
254 “enhanced spatial selection” strongly impacts the dispersal trait because of the gradient in that
255 trait along the invasion front driven by spatial sorting. The distributions of the life-history traits,
256 on the other hand, show no such gradient (because these phenotypes are not associated with
257 dispersal distance) and so do not receive the additional evolutionary boost afforded to dispersal.

258 Finally, the scenarios we examined allow us to partition the spread increase in L_+D_+ relative
259 to L_0D_0 into contributions from natural selection on life-history traits (24%), spatial selection
260 on the dispersal trait (39%), and enhanced spatial selection on the dispersal trait that occurred
261 as a byproduct of the evolution of the life-history traits and their impact on population growth
262 (37%). Thus, changes in the dispersal trait were directly responsible for most (76%) of the
263 increase in spread due to evolution, but nearly half of that increase was an indirect result of
264 life-history evolution. The acceleration of cane toad spread in our model therefore appears not
265 to have been dominated by any single trait or evolutionary process.

266 DISCUSSION

267 Although previous work has demonstrated genetically based evolution of cane toads (Phillips
268 *et al.*, 2006, 2008b, 2010a; Phillips, 2009) and suggested that these changes might have ac-
269 counted for accelerating spread across northern Australia (Phillips *et al.*, 2006, 2008a), ours is

270 the first study to connect these disparate data on phenotype change and spatial spread with a
271 mathematical model. Our results show that modest heritabilities (0.10-0.21) are sufficient to
272 account for phenotype changes between the introduction site and the invasion front (Fig. 2)
273 and that these changes have large enough effects on population growth and dispersal to have
274 a substantial impact on spread on the ecological time scale of interest (Fig. 3). By comparing
275 phenotype changes and distances spread under different scenarios about the initial heritability
276 of life-history and dispersal traits, we found that no single evolving trait dominated spread dy-
277 namics. Rather, life-history evolution, dispersal evolution, and an interaction between the two
278 all appear to have made important contributions to the spatial spread of this invasive species. In
279 particular though, the increase in spread due to an interaction between life-history and disper-
280 sal evolution identifies a new mechanism, which we term “enhanced spatial selection,” for trait
281 change during spread and highlights the importance of this process for the spread of invasive
282 species.

283 The process of enhanced spatial selection would seem to be a general outcome of allowing
284 both life history and dispersal to evolve during spread. If we recall that spatial selection on
285 dispersal is the interaction between spatial assortment of dispersal phenotypes and differential
286 population growth driven by density release, then we can see that higher population growth
287 resulting from natural selection should indeed enhance spatial selection. More generally, any
288 factor that modulates a gradient in population growth in the direction of spread has the potential
289 to modulate dispersal evolution, enhancing or suppressing it depending on the direction of the
290 gradient. For a species spreading into increasingly inhospitable conditions (e.g., to track a
291 shifting climate), the action of spatial selection could be suppressed. Indeed, even without an
292 environmental gradient, spatial selection can be suppressed by Allee effects (Travis & Dytham,
293 2002). Together, however, our model and the body of evidence for cane toads suggest that
294 natural selection on life-history traits enhanced the evolution of dispersal traits in the cane toad
295 invasion of northern Australia.

296 Moreover, the various mechanisms for trait change during range expansion are not simply

297 evolutionary curiosities but also drivers of ecological dynamics. In the case of the cane toad
298 invasion of northern Australia, our analysis posits that genetically based trait change could have
299 led to a more than twofold increase in the distance spread by cane toads over 72 years. This
300 impact on spread appears equally attributable to three distinct mechanisms: natural selection,
301 spatial selection, and enhanced spatial selection. In other systems, the contributions of these
302 mechanisms will likely differ from those seen here, depending on patterns of genetic variation
303 and selection. In related but strictly ecological analyses, some have shown that differences in
304 demographic characteristics contribute more to differences in spread rates (Caswell *et al.*, 2003;
305 Smith *et al.*, 2009), whereas others have shown that differences in dispersal play a larger role
306 (Caswell *et al.*, 2003; Bullock *et al.*, 2008; Jongejans *et al.*, 2008). Although we are unable to
307 provide general answers about when certain ecological or evolutionary mechanisms contribute
308 more or less to spread, we do make the important advance of demonstrating the impact that
309 rapid evolution can have on spread in a natural system and discovering the importance of an
310 interaction between life-history and dispersal evolution. By building on our modeling framework
311 (i.e., applying it to other species or developing novel extensions of it), others will be able to
312 provide additional clarity on these issues.

313 Although our model is well-supported by empirical studies and well-suited to assessing the
314 relative impacts of different evolutionary scenarios on spread, it should not be misconstrued
315 as an attempt to thoroughly recreate the spread dynamics of cane toads. Most notably, the
316 distance spread in our most plausible evolutionary scenario was only 58% of the empirical es-
317 timate of distance spread by Urban *et al.* (2008). This is unsurprising for several reasons, of
318 which we comment on a few. First, and perhaps most importantly, spatial spread is an in-
319 herently stochastic process that is extremely difficult to predict, even in controlled laboratory
320 environments (Melbourne & Hastings, 2009). Second, any number of factors not considered
321 by our model could have affected spread, including spatiotemporal variability (Grosholz, 1996),
322 genotype-by-environment interactions (Bowler & Benton, 2005), and subtle environmental gradi-
323 ents (Urban *et al.*, 2008). Incorporating such factors would likely only affect absolute predictions

324 of the model rather than relative predictions across evolutionary scenarios, which are our focus.
325 Third, more exhaustive sampling of daily displacements by Phillips *et al.* (2008a) or assuming
326 a leptokurtic form of the individual dispersal kernel could have led to much greater estimates
327 of distance spread, but doing so would not be supported by available evidence and would not
328 impact our findings. Despite these limitations, our analysis is clear about the relative impacts
329 of different evolutionary mechanisms on spread, and it yields quantitative results about distance
330 spread on the same order of magnitude as empirical estimates.

331 Our analysis should also not be construed as a complete recreation of the evolutionary history
332 of cane toads in Australia. One source of uncertainty is that, for lack of historical evidence, we
333 initialize phenotypes in the model based on contemporary estimates from Gordonvale and thus
334 assume that the population there has evolved very little. The behavior of the model at the
335 introduction site contradicts this assumption, however. With all $h_i^2 > 0$, changes in \bar{z}_T , \bar{z}_M ,
336 and \bar{z}_D at the introduction site equal 76%, 84%, and 4% of the changes at the invasion front.
337 This assumption of little change at the introduction site is therefore appropriate for dispersal
338 evolution but less so for life-history evolution. Rather than bend the model to accommodate this
339 behavior by calibrating initial phenotype distributions and heritabilities to potentially unrealistic
340 values, we find a more parsimonious reconciliation of this discrepancy to be that the model of
341 life-history evolution is simply less appropriate for the high-density context of an established
342 population than it is for a low-density population on the invasion front (Phillips *et al.*, 2010b).
343 Although further empirical studies of life-history evolution in the wild could provide clarity
344 on this issue, such refinements would have little bearing on our qualitative results about the
345 impact of different evolutionary forces on the dynamics of frontal populations. Another source
346 of uncertainty is the evolutionary impact of a secondary introduction sometime around 1964-
347 1968 in an area about 150 km ahead of the invasion front (Estoup *et al.*, 2004). Joint analysis
348 of genetic data and historical distribution records, however, indicates relatively little signal of
349 this secondary introduction in the genetics of descendent populations (Estoup *et al.*, 2010).
350 Incorporating this secondary introduction into our model may therefore have little impact on

351 our estimates of evolution along the invasion front but could boost our estimate of distance
352 spread.

353 Perhaps the most serious limiting assumption of our model is that about genetic variation.
354 Our model does allow for changes in genetic variation due to selection and gene flow and for
355 departures of the genotype distributions from normality (Turelli & Barton, 1994). It does not,
356 however, account for changes in genetic variation due to drift. Even so, because we calibrated
357 trait heritabilities to achieve empirically estimated shifts in trait means, allowing for a decay
358 in heritability over time due to drift would have simply required a recalibration of our initial
359 heritability estimates. The main consequence of this is that the initial heritability values that
360 we used could somewhat underrepresent the true heritabilities during the early stages of the
361 cane toad's invasion. In addition to genetic variation within traits, our modeling framework
362 is also capable of accommodating genetic covariances among traits, but we have no such data
363 for cane toads and likely will not for some time due to the difficulty of fully estimating the \mathbf{G}
364 matrix for non-model organisms. Despite these uncertainties, our model accords with empirical
365 demonstrations that life-history and dispersal traits in cane toads are heritable and that they
366 have shifted during the cane toad's rapid range expansion across northern Australia (Phillips
367 *et al.*, 2006, 2008a, 2010a; Phillips, 2009).

368 CONCLUSION

369 Overall, we have developed a useful framework with which to explore the role of evolution
370 in the ecological dynamics of range expansion. Applying this framework to cane toads, our
371 results indicate that rapid evolution of life-history and dispersal traits at the invasion front
372 could have led to a more than twofold increase in the distance spread by cane toads across
373 northern Australia. Additionally, by partitioning evolutionary impacts on spread into those
374 acting separately on life-history and dispersal traits, our analysis reveals a new mechanism for
375 dispersal evolution under range shift: enhanced spatial selection. These results speak broadly
376 to the importance of incorporating the capacity for trait change into spread models and into

377 ecological models more generally. Doing so can have dramatic impacts on the predictions flowing
378 from such models and thereby on our capacity to anticipate and manage the dynamics of species
379 undergoing range shifts during invasion or in response to climate change.

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505 FIGURES

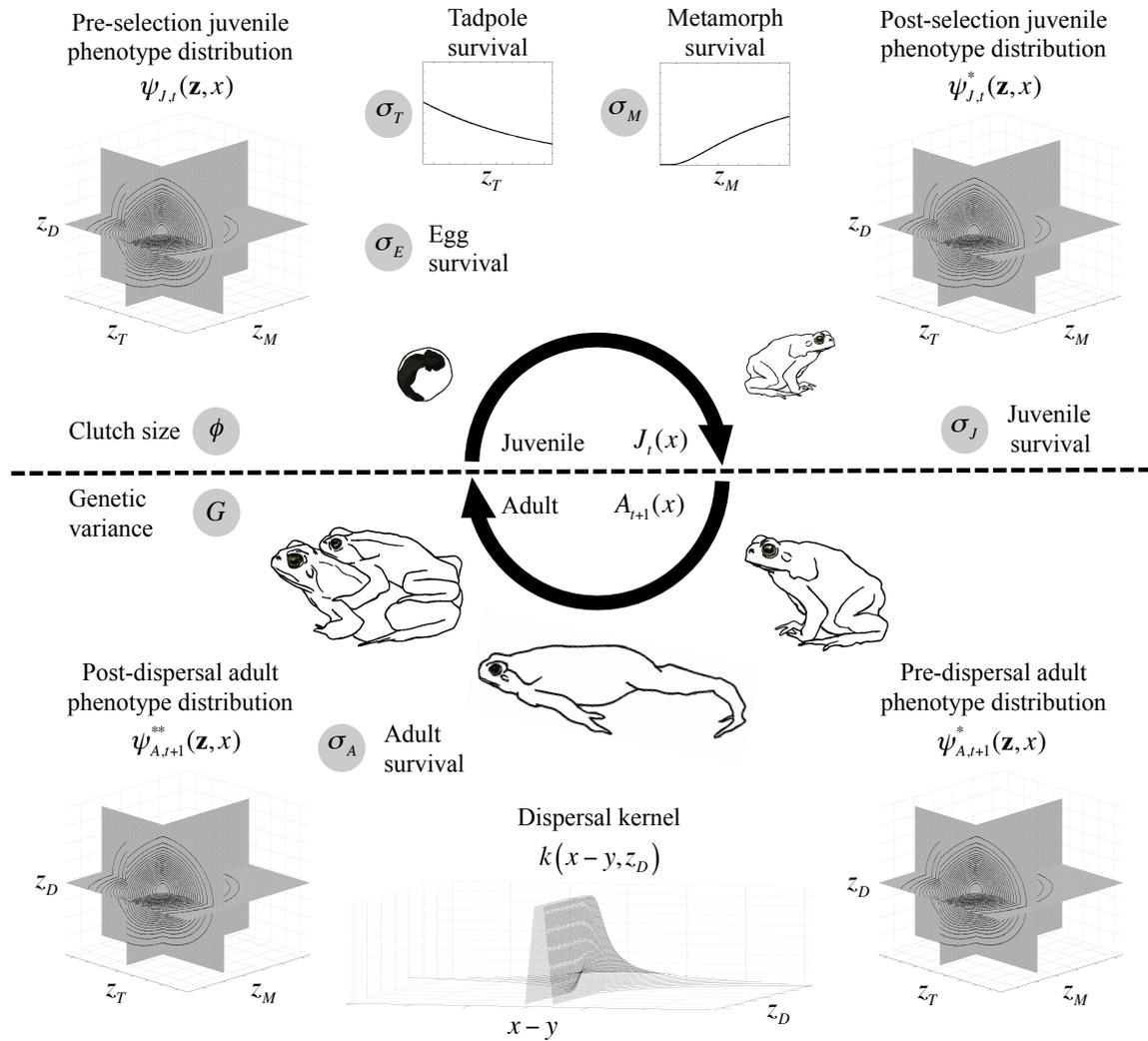


Figure 1. Model schematic showing the progression of the cane toad life cycle (clockwise) through the juvenile (top) and adult (bottom) stages. The four slice plots of the $\mathbf{z} = (z_T, z_M, z_D)$ phenotype space at some point in space x show the distinct phenotype distributions tracked by our model: $\psi_{J,t}(\mathbf{z}, x)$ (top left) is that of a virgin cohort; $\psi_{J,t}^*(\mathbf{z}, x)$ (top right) results from phenotype-dependent tadpole and metamorph survival; $\psi_{A,t+1}^*(\mathbf{z}, x)$ (bottom right) is comprised of newly matured adults and those surviving from past years; $\psi_{A,t+1}^{**}(\mathbf{z}, x)$ (bottom left) is the result of phenotype-dependent dispersal from and to each location. Model parameters are highlighted with gray backgrounds.

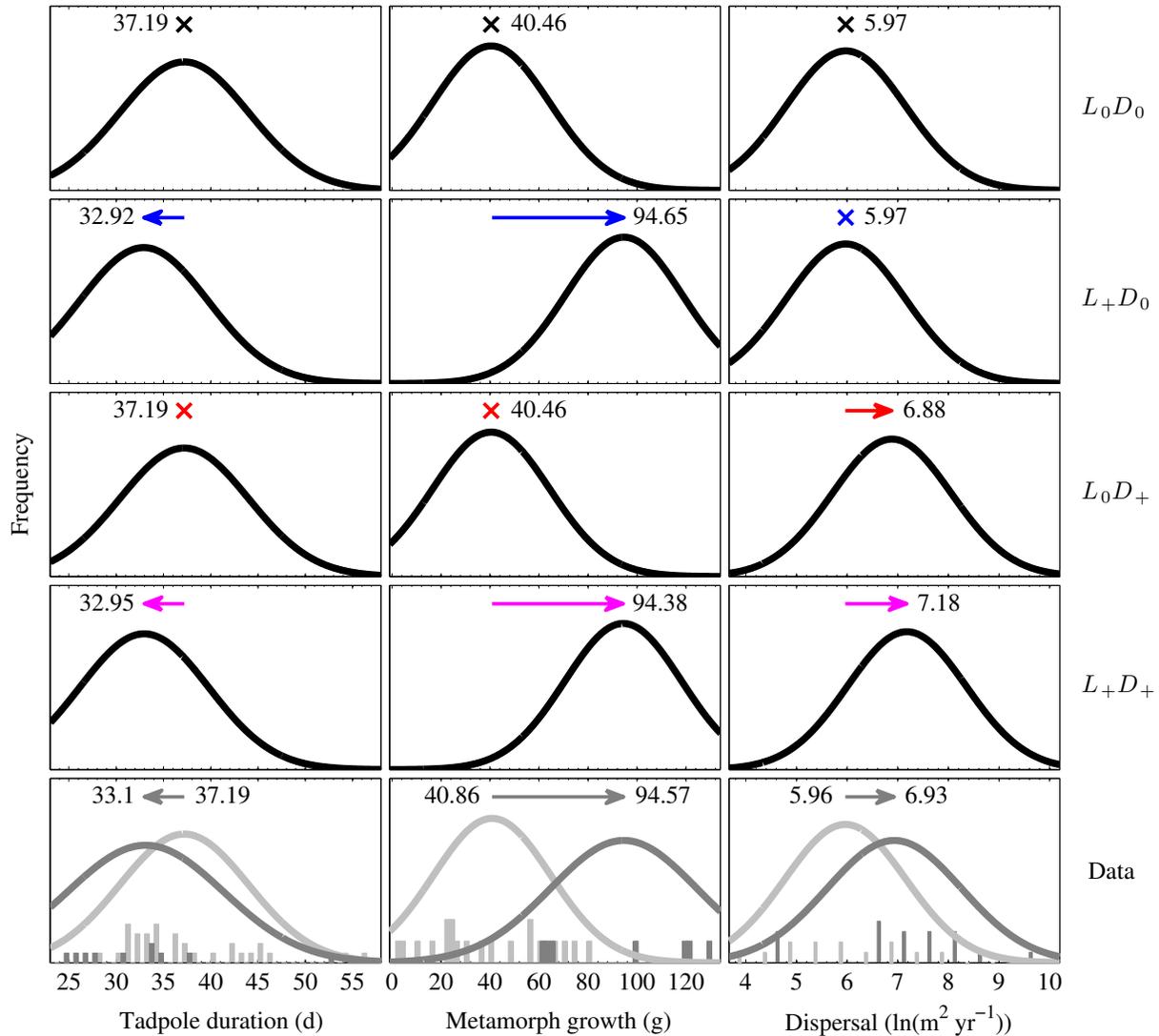


Figure 2. Phenotype marginal distributions of tadpole duration (left column), metamorph growth (center column), and dispersal tendency (right column), estimated from empirical data from Gordonvale (bottom row, light gray) and Timber Creek (bottom row, dark gray) and calculated with our model under different scenarios (top four rows) about initial heritability of life-history (L) and dispersal (D) traits (0, not heritable; +, heritable). Arrows show direction and magnitude of difference in phenotype means between Gordonvale and Timber Creek, and their colors correspond to the scenarios in Fig. 3. A × indicates no change in phenotype mean.

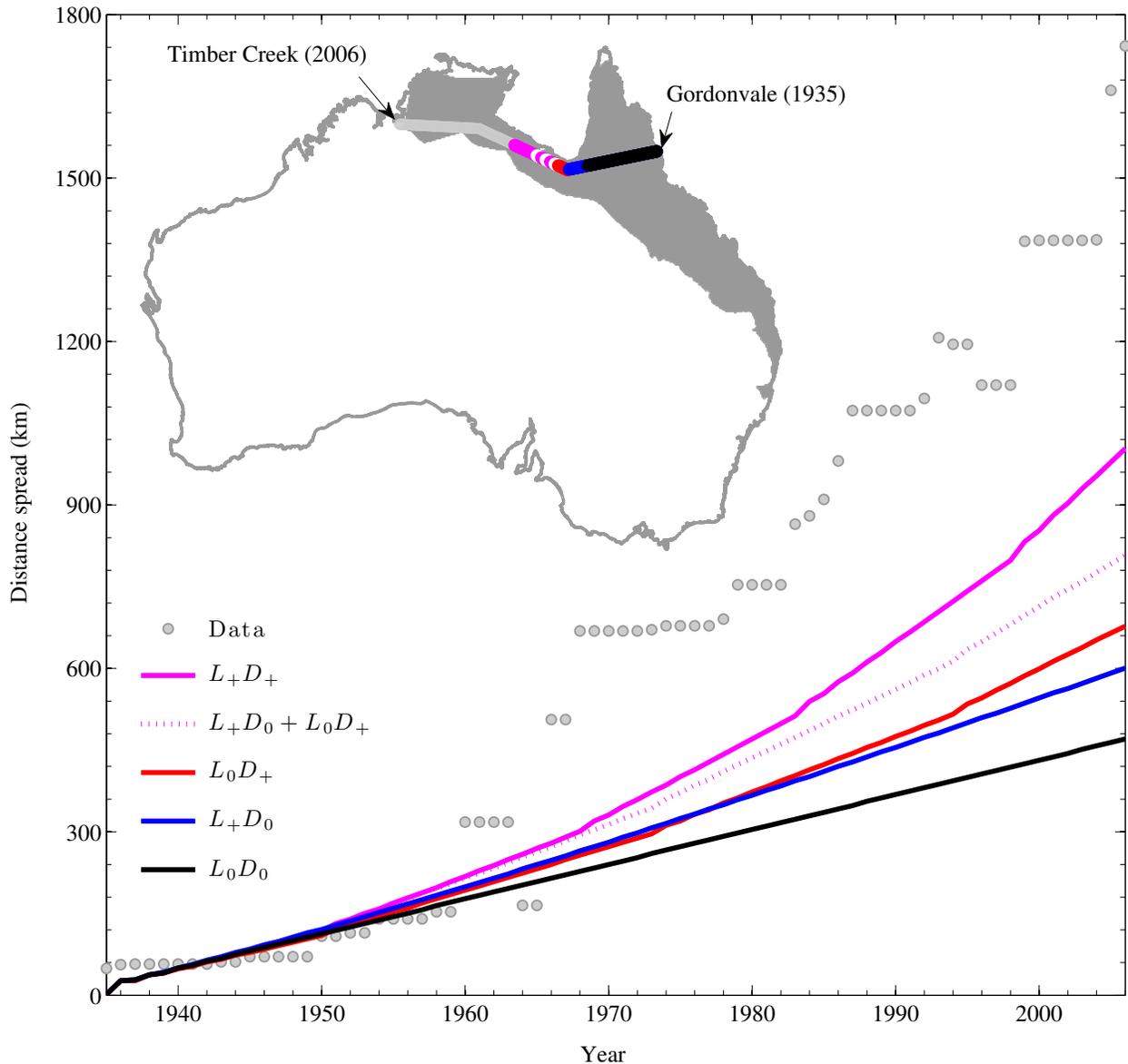


Figure 3. Distance spread by cane toads, modeled under different evolutionary scenarios. Solid lines correspond to model scenarios about initial heritabilities (0 or +) of life-history (L) and dispersal (D) traits. The dashed line shows the sum of the increases in distance spread in scenarios in which only one type of trait evolved (L_+D_0 and L_0D_+) relative to when neither evolved (L_0D_0), which contrasts with the scenario in which both traits evolved simultaneously (L_+D_+). The line between Gordonvale and Timber Creek shows the one-dimensional path along which spread was modeled, and colors on that line show distance spread by 2006 under the different model scenarios. Although the model was neither fit to spread data nor intended to fully recreate empirical patterns, we show an empirical estimate of cane toads' distance spread over time and their range in Australia as of 2006 for context (from Urban *et al.*, 2008).